

Response of Penaeid Prawns (*Metapanaeus monoceros*) to Textile Dye Industrial Effluents (TDIE): An Indicator of Stress

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Abstract

Studies on acute toxicity of textile effluents, collected from various Textile Dye Industries from Erode industrial area, Tamil Nadu, have been carried out. The juveniles of pink prawns (*Metapanaeus monoceros*) were exposed to a mixture of effluent samples in five different concentrations prepared in seawater, using static acute toxicity test protocol to obtain the 96-h LC₅₀. The five different dilution factors of effluents made up with seawater were 200, 100, 50, 33.33, and 20 respectively. The environmental factors such as salinity, temperature, dissolved oxygen and pH were monitored from different effluent dilutions every 24 hours during the experiment till 96 hours. Each tank of different effluent dilution contained 10 test organisms in 3L of effluent content besides a control tank. The results showed that the 96-h LC₅₀ of Erode effluent samples for prawns is 29 ppm, the upper limit is 41 ppm and the lower limit is 17 ppm. Maximum mortalities occurred at a dilution factor 20 i.e. concentration of 50 ppm. This indicated a greater toxicity of these tested textile effluents from Erode (with a Dilution Factor of 20) as compared to the effluents tested earlier by other workers from different dye industries in Erode area (with a Dilution Factor of 5).

Keywords: Textile industries effluent; Erode; Acute toxicity; LC₅₀; Penaeid prawns

Introduction

Erode district in Tamil Nadu, India is a rapidly growing city and is one of the well known center for textile industries. It has about 736 big textile dyeing units and several small dying units. All these dyeing units utilize large quantity of water for their process and generate huge quantity of wastewater or effluents [1], and the effluents are normally discharged in the adjacent riverine environments or in small streams in the Erode area. Such practice was going on over a period of last few decades and it has seriously affected the aquatic environment. The textile effluents are important sources of toxic substances, organic and inorganic in nature. Considering the gravity of situation of the ongoing environmental degradation in Erode, the Textile Ministry, Govt. of Tamil Nadu, with the support of the Department of Science and Technology (DST), New Delhi proposed to have a common effluent pipeline to discharge the effluents from all the textile dyeing industries from Erode directly into the coastal sea after treatment. Such a huge discharge of effluents can have a devastating effect on the coastal marine water quality and productivity [2]. The evaluation of toxicity of these effluents prior to discharge is therefore highly necessary for understanding the extent of its toxicity and the measures to be taken for reducing the toxicity before it is discharged into the sea. In order to understand the toxicity limit of a mixture of textile dye industrial effluent (TDIE), the effluents, particularly the RO (reverse osmosis) rejects were collected from each of the 9 selected dye industries in Erode area during May 2011 and after mixing in equal proportions were used for toxicity studies using static acute toxicity test. This unique study is maiden attempt to explore the toxicity of effluents when released in the marine environment of the east coast of India.

Material and Method

The TDIE or the RO rejects were collected from 9 textile industries from the Erode Industrial area. Part of these collected effluents was used for physico-chemical analyses at the site itself and the other part was taken to the National Institute of Oceanography laboratory at Goa for bioassays studies.

The parameters analyzed for physico-chemical studies of the effluents included pH, turbidity, salinity, phosphate, nitrite, nitrate, ammonia, silicate, urea as per Strickland and Parsons, Grasshoff and APHA [3-5]. Total suspended solids (TSS) were measured by filtration using Filtration pump (Aspirator Vacuum). Millipore vacuum filter pump was used for the filtration and the samples were filtered through pre-weighed 0.45 µm filter papers. TSS was calculated from the initial and final weights of the filter paper after filtration. The heavy metals in the effluents were analyzed using ICP-AES. Trace metals were analyzed by Ion chromatography (DIONEX ICS-2500), while the mercury was analyzed by Mercury analyzer, (BUCK Scientific 410 Mercury analyzer). The total organic carbon analyses were carried out using TOC analyzer (SHIMADZU TOC-VCPH).

For bioassay study, all the effluent samples were mixed together in equal proportion and a known quantity of the mixed sample was used for toxicity test after dilution. A known volume of sea water was measured using a measuring cylinder into a clean, dry bioassay tank and a pre determined volume of the mixed effluent was added to the sea water to make it up to 1000ml (total volume of test media) to achieve the desired test concentration. The toxicity test was done on prawns (*Metapanaeus monoceros*), using each of the diluted effluents having different dilutions of 200, 100, 50, 33.33 and 20, along with a control with no effluent in it. The bioassay test used was the acute static non renewal toxicity test following Sridevi and Kanmani [6], to understand

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the response of Penaeid prawns (*Metapanaeus monoceros*) to each of the diluted textile dye industrial effluent (TDIE) for evaluating toxicity.

Test animal

The Penaeid prawns, *M. monoceros*, an economically important crustacean species are harvested in considerable quantities from estuarine mangrove environments and traditional aquaculture ponds of Goa along the west coast of India, with riverine/estuarine environments [7]. This species is in great demand due to its high nutritional value and forms a cheap source of animal protein for lower and middle income groups. The prawns *Metapanaeus monoceros* (Arthropoda, Crustacean, Decapoda) of similar sizes ranging from 5 to 7cm with a body weight of 1.5 to 2 g were utilized for the bioassay in this study. These prawns were regularly obtained from Britona, Goa and efforts were made to obtain the prawns from the same location throughout the experiment period to reduce variability in biotypes. These prawns were selected for the toxicity study because they are very sensitive to contaminants and show mortality at lethal concentrations.

The penaeid prawns were kept in a tank filled with 40 L water from the estuary and left for a minimum of 7 days to allow acclimatization of prawns to experimental conditions (Temp.: $28 \pm 1^\circ\text{C}$; pH 7.98 to 8.10, Salinity: $12 \pm 2\text{‰}$). During the period of acclimatization, the prawns were fed on the prepared fish meal (5% of the average body weight of total animals) (Table 1). The tank was continuously aerated with an air pump. Dead shrimps, if any, were immediately removed from the acclimatization tank. The 50% of water was changed daily to prevent the accumulation of waste metabolites and decaying food materials [8]. In order to reduce the excreted waste in the test tanks, the feeding was stopped 24hrs prior to the commencement of acute bioassay tests.

Test method

Short-term LC_{50} (median lethal concentration) toxicity tests were carried out according to the methods described by the APHA [5]. The prawns were sampled randomly from the holding tanks and transferred to the test and control solutions. The toxicity testing was done following the acute static non renewal method [6]. Toxicity testing is one of the common approaches for studying the impact of contaminant on marine organisms.

Tests are to be static tests. They should conform to the normal requirements of bioassay procedures and in particular, there should be:-

- At least one control chamber and four test chambers each, having a different concentration of test effluent.
- Full randomization of chamber and animals.
- Test chambers are to be gently aerated to prevent stratification of test media. At least ten animals per test chamber but the biomass should not exceed one gram per liter and dissolved oxygen should not fall below 4 mg/l in 24 hrs.

These tests provide an initial hazard assessment. The toxicity of the contaminants was assessed using the routine acute static bioassay method based on the determination of LC_{50} values for 24, 48, 72 and 96hrs exposure period. The median lethal concentration (96-h LC_{50}) and 95% confidence intervals were determined with a computer-based program described by Finney [9]. Test data were reported as the toxicity response curves showing the Lethal Time (LT) value against each concentration tested. The 96 hr LC_{50} values were determined from the response curves.

Assessment of quantal response (mortality)

Prawns, *M. monoceros*, were assumed to be dead if there was no movement of the appendages, opercula, or mouth even when prodded with a blunt glass rod.

Bioassay

The prawns were exposed to effluent sample mixtures with varying dilution factors of 200, 100, 50, 33.33, and 20, which corresponds to the concentrations of 5 ppm, 10 ppm, 20 ppm, 30 ppm and 50 ppm respectively and an untreated control with no effluent. Mortality assessment was carried out once in every 24 h over a period of 4 days.

Statistical analyses

Toxicological dose-response data involving quantal response (mortality) studies were analyzed by probit analysis [9]. The indices of toxicity measurement derived from these analyses were LC_{50} (median lethal concentration that causes 50% response (mortality) of exposed organisms), LC_{95} (lethal concentration that causes 95% response (mortality) of exposed organisms), LC_{05} (lethal concentration that causes 5% response (mortality) of exposed organisms) and TF (toxicity factor for relative potency measurements e.g., ratio of 96 h LC_{50} of a compound to LC_{50} values at equivalent time intervals).

Results and Discussion

Physico-chemical parameters

The physico-chemical characteristics of the RO reject samples from the textile units showed that the ammonia and nitrate concentrations are very high, to the extent of 3048.94 $\mu\text{mol/L}$ and 931.47 $\mu\text{mol/L}$ respectively. Such high concentrations of ammonia and nitrate are found to act toxic to the central nervous system (CNS) of aquatic fishes, molluscs and crustaceans [10]. So, the higher concentrations of ammonia and nitrate present in the RO rejects can cause great damage to the aquatic life in the area if the reject is discharged in the sea.

Normally, the phosphate and nitrate are the two main nutrients responsible for biological productivity in the sea [11]. However, Chen et al. [12], during their experiments found that the ammonia is more toxic than nitrite, and nitrate is slightly toxic to shrimp. The median lethal concentrations (LC_{50}) of ammonia and nitrite have been estimated for penaeid shrimp post larvae, such as *Penaeus monodon*, *P. chinensis*, *P. puulensis*, and *P. japonicus* by many workers viz. Chin and Chen, Chen and Chin, Chen and Lin [13-15]. They showed that ammonia causes damage to their central nervous system. In case of penaeid shrimps, the presence of high concentrations of ammonia in the environment may affect their acid-base balance, haemolymph, osmolarity, nitrogen metabolism, respiration, and growth rates and may also enhance molting [16]. Studies on the effect of nitrite on freshwater animals by Lewis and Morris [17], indicated the inducement of reversible methaemoglobin formation, which is an inability to transport oxygen to the tissues [10].

Besides nutrients, very high levels of TDS upto 39,000 mg/L also have been observed in these RO reject samples. Boyd and Claude [18], have shown that most of the aquatic ecosystems involving mixed fish fauna can tolerate TDS levels of upto 1000 mg/L. Comparatively, the observed high TDS (av.19617.3 mg/l) in Erode effluent mixture can have a significant negative effect on fisheries of the region.

Wet effluent toxicity test

The LD_{50} represents a dose that would kill half of the population that

is exposed to it. "WET" is a term used to describe the adverse effects or toxicity to a population of aquatic organisms caused by exposure to an effluent. This toxicity is experimentally determined in the laboratory by exposing sensitive organisms (usually surrogate organisms representative of those found in the environment) to effluents using WET tests. Responses assessed usually include survival, growth, and/or reproduction. WET testing is used to assess and regulate the combined effects of all the constituents of a complex effluent rather than the conventional methods of controlling the toxicity of single chemical or constituent. Whole effluent toxicity (WET) tests can be more realistic than the study of individual pollutants and are recognized as practical and effective tools for the assessment of combined effects of toxic substances on aquatic ecosystems. Though WET tests do not provide indication of the specific cause(s) of toxicity, they contribute to overall effects that a mixture of pollutants might produce on the aquatic environment [19]. In order to know the level of toxicity of RO reject samples mixture to the aquatic life, the prawns were exposed to a mixture of effluent samples diluted to varying concentrations (dilution factors of 200, 100, 50, 33.33, and 20). Prior to the exposure of prawns, they were found to be active but after exposure to the industrial effluent mixtures of varying concentrations, their activity started showing an increase as seen from their agitated swimming. Thereafter, the activity showed a gradual reduction with a loss of equilibrium, progressive weakness and periods of inactivity and quiescence. Eventually, the organisms showed mortality, and this was confirmed when there was no response to gentle prodding to a blunt pair of forceps. Following the death, the exoskeleton of the prawns turned whitish from their original transparent color. The intensity of these symptoms was directly related to the concentration of effluent as well as the duration of exposure. As seen from Figure 1, a zero percent mortality was observed in effluent with 200 dilution factor, whereas 70% mortality was observed in effluent with 20 dilution factor after 96 hours period.

This showed that as the dilution factor increases the mortality percent decreases. In toxicity studies carried out by Sridevi and Kanmani [6], a higher mortality rate was observed in an effluent with a dilution factor of 5 using ionized textile dyeing industries waste water. The present study carried out on the toxicity evaluation of RO reject effluents from Erode textile industries showed that the toxicity levels are not meeting the standards as recommended by the MINAS (Minimum National Standards as proposed by the Central Pollution Control Board involving three other laboratories i.e. Gujarat Pollution Control Board, Gandhinagar; National Institute of Occupational Health (NIOH), Ahmedabad and National Environmental Engineering Research Institute, Nagpur) for dye and dye intermediate industries. MINAS has recommended a toxicity factor of 4.0 (Dilution factor 4) for treated effluents from dye and dye intermediate industries. This implies that the treated industrial effluent should not show acute toxicity for fish, if diluted four times. Comparatively, our observations show that the toxicity of the mixture of effluent samples from Erode is high, suggesting that they are more toxic even at a dilution factor of 20. This is also higher as compared to the toxicity test done by Sridevi and Kanmani [6], using effluents from other textile dyeing industrial units in Erode area.

The RO reject samples from all the stations were made into one mixture and the dilutions were made by using seawater and a control without the effluent. To express the lethal concentration of the effluents with dilution factors of 200, 100, 50, 33.33, 20, the dilution factors were converted to ppm. The penaeid prawns were kept in the tanks containing effluents with different dilution factors after the acclimatization period. It was observed that after 24hours there was no mortality in all the

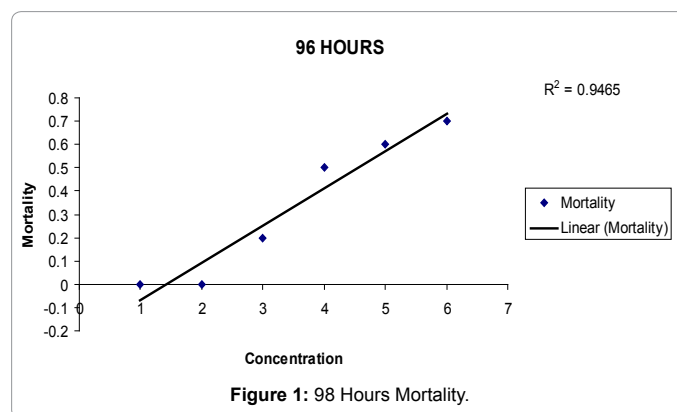
concentrations. This indicated that there was no effect on the penaeid prawns for the dilution factors of 200, 100, 50, 33.33 and 20 for an exposure period of 24 hours (Table 2). Thereafter, at 48 hours, the observations showed 10% mortality in both the effluents with 33.33 and 20 dilution factors. Rest of the diluted effluents showed zero mortality after 48hours (Table 3). Then, after 72 hours there was an increase in mortality from 10% to 40% in both the effluents with dilution factors of 33.33 and 20. 30% mortality was observed in effluent with 50 dilution factors and 10% mortality was observed in effluent with 100 dilution factor whereas, no mortality was observed in effluent with 200 dilution factor (Table 4). After 96 hours, 70% mortality was observed in effluent with 20 dilution factor; 60% mortality in 33.33 dilution factor; 50% mortality in 50 dilution factor and 20% mortality in 100 dilution factor (Table 5). Throughout the experiment there was no mortality observed in the control tank and in the effluent with 200 dilution factor. This clearly indicates that as the dilution factor decreases, the mortality of prawns' increases with time. The effluent with a Dilution Factor of 200 can be considered as No Observable Effect Level (NOEL) as there was no mortality observed in prawns even after 96 hours of exposure.

Probit analysis

Probit Analysis is a method of analyzing the relationship between a stimulus (dose) and the quantal (all or nothing) response. In a typical quantal response experiment, groups of animals are given different doses of an effluent or a chemical and the percent dying at each dose level is recorded and the data analyzed using Probit Analysis.

The effect of any substance depends on a number of factors. The most important factor is the dose-time relationship (Table 6). The dose-time relationship indicates the quantity of substance involved and how often the exposure to the substance has occurred. This relationship gives rise to two different types of toxicity studies, and they are acute and chronic toxicity studies.

The Probit model assumes that the percent response is related to the logarithmic dose as the cumulative normal distribution. That is, the log doses may be used as variables to read the percent dying from the cumulative normal (Table 7). Using the normal distribution, rather



Mean wet weight (g)	Daily feeding rate (% of body weight)
<5	0
5-15	7
15-25	5
>25	3

As fed weight of diet/wet biomass of prawns x 100

Table 1: Weight dependent feeding rates of *M. monoceros*.

Sr. No.	Dilution Factor	Concentration (ppm)	Total animals	Mortality	Mortality percentage
1	0	0	10	0	0%
2	200	5	10	0	0%
3	100	10	10	0	0%
4	50	20	10	0	0%
5	33.33	30	10	0	0%
6	20	50	10	0	0%

After 24 hours exposure to the five different concentrations of the effluent sample there was no mortality observed

Table 2: 24 Hours Mortality.

Sr. No.	Dilution Factor	Concentration (ppm)	Total animals	Mortality	Mortality percentage
1	0	0	10	0	0%
2	200	5	10	0	0%
3	100	10	10	0	0%
4	50	20	10	0	0%
5	33.33	30	10	1	10%
6	20	50	10	1	10%

After 48 hours exposure period it was observed that there is 10% mortality in 30 ppm and 50 ppm concentrations of effluent sample

Table 3: 48 Hours Mortality.

Sr. No.	Dilution Factor	Concentration (ppm)	Total animals	Mortality	Mortality percentage
1	0	0	10	0	0%
2	200	5	10	0	0%
3	100	10	10	1	10%
4	50	20	10	3	30%
5	33.33	30	10	4	40%
6	20	50	10	4	40%

After 72 hours exposure period it was observed that in 10ppm concentration of effluent there was 10% mortality, for 20ppm there was 30% mortality, 30ppm there was 40% mortality and for 50 ppm there was 40% mortality observed

Table 4: 72 Hours Mortality.

Sr. No.	Dilution Factor	Concentration (ppm)	Total animals	Mortality	Mortality percentage
1	0	0	10	0	0%
2	200	5	10	0	0%
3	100	10	10	2	20%
4	50	20	10	5	50%
5	33.33	30	10	6	60%
6	20	50	10	7	70%

After 96 hours exposure to the effluent sample, in 10ppm concentration there was 20% mortality, in 20ppm there was 50% mortality, in 30ppm concentration there was 60% mortality and in 50ppm concentration there was 70% mortality

Table 5: 96 Hours Mortality.

Dilution Factor	Concentration (ppm)	Percentage Mortality			
		24 hours	48 hours	72 hours	96 hours
0	0	0	0	0	0
200	5	0	0	0	0
100	10	0	0	10	20
50	20	0	0	30	50
33.33	30	0	10	40	60
20	50	0	10	40	70

Table 6: Percentage mortality for 24 hrs, 48 hrs, 72 hrs, and 96 hrs for effluent.

than other probability distributions, influences the predicted response rate at the high and low ends of possible doses, but has little influence near the middle. Hence, much of the comparison of different drugs is done using response rates of fifty percent. The probit model may be expressed mathematically as follows: $P = \alpha + \beta [\log_{10}(\text{Dose})]$

Where, P is five plus the inverse normal transform of the response rate (called the Probit). The five is added to reduce the possibility of

negative probits, a situation that caused confusion when solving the problem by hand. When using Finney method, the regression line is created for logs of doses. Some programs in Finney method adjust 0% and 100% lethality under the formula $49\%/N$, instead of $25\%/N$.

The median lethal concentration (96-h LC_{50}) and 95% confidence intervals were determined as described by Finney [9]. As there was no mortality observed after 24 hours, it is considered as a zero value

for the LC₅₀. Then after 48 hours the LC₅₀ was observed as 119.6 ppm with 158.52 ppm as the upper limit and 80.7 ppm as the lower limit (Table 8 and Figure 2). The intercept for 48hours LC₅₀ value is 2.9629 and Chi square value is 31.839 Figure1. After 72 hours, the LC₅₀ value is 51 ppm with 68.69 ppm as the upper limit and 33.3 ppm as the lower limit (Table 9 and Figure 3).. The intercept for 72 hours, the LC₅₀ value is 3.6959 and the Chi square value is 0.5366. After 96 hours exposure period, the LC₅₀ value was 29.17 ppm with an upper limit of 41.4 ppm and the lower limit of 16.9 ppm Figure 3. The intercept for 96 hours LC₅₀ value is 3.9211 and Chi square value is 0.4088 (Table 10 and Figure 4). The 96-h LC₅₀ values for other shrimp species ranged between 23.7 ppm for *P. semisulcatus* and 70.9 ppm for *P. monodon* [12].

Conclusion

In the present study, the effluent samples (RO rejects) were obtained

from 9 dye industries located at Erode in Tamil Nadu and used for toxicity testing of prawns (*Metapanaeus monoceros*) collected from off Britona, Goa. All the 9 effluents were mixed together and a mixture of the effluents from Erode textile dye industries was subsequently diluted further with different dilution factors of 200, 100, 50, 33.33 and 20 and exposed to prawns along with a control with no effluent for evaluating toxicity. Penaeid prawns (*Metapanaeus monoceros*) are widely distributed in different size groups. They are commercially important cultivable species and have high export demand in India. This species was used for acute toxicity testing because of their large availability, sensitivity and relative ease of handling.

The experiment showed that the Lethal Concentration for killing 50% organisms (LC₅₀) increases with a decrease in dilution and an increase in time. A high mortality (70%) was observed in effluent

Exposure period	LC ₅₀	Upper Limit	Lower Limit	Intercept	Chi square
24 h	-	-	-	-	-
48 h	119.6 ppm	158.52ppm	80.7 ppm	2.9629	31.839
72 h	51ppm	68.69ppm	33.3ppm	3.6959	0.5366
96 h	29.17ppm	41.40ppm	16.9ppm	3.9211	0.4088

Table 7: Median Lethal concentration (LC₅₀) (with 95% confidence intervals) for TDIE.

Alpha value (for confidence interval)	0.05						
Probit Analysis - Finney Method [Lognormal Distribution]							
Log10[Dose (Stimulus)]	Actual Percent (%)	Probit Percent(%)	N	R	E(R)	Difference	Chi-square
0.699	0.025	0.0002	10	0.25	0.0021	0.2479	29.2063
1.	0.025	0.0025	10	0.25	0.0253	0.2247	2.
1.301	0.025	0.0187	10	0.25	0.1873	0.0627	0.021
1.4771	0.1	0.0487	10	1.	0.4867	0.5133	0.5414
1.699	0.1	0.1303	10	1.	1.3027	-0.3027	0.0704
Chi-square							
Chi-square	31.839						
Degrees Of Freedom	3						
p-level	0.						
Dose (Stimulus) Percentile							
Percentile	Probit (Y)	Log10[Dose (Stimulus)]	Standard Error	Dose (Stimulus)	Standard Error		
1	2.6732	1.1986	-0.1425	15.7981	-5.2755		
5	3.3548	1.4824	-0.1189	30.3671	-8.4199		
10	3.7183	1.6338	-0.1149	43.028	-11.5156		
16	4.0056	1.7534	-0.0543	56.6736	-7.1028		
20	4.1585	1.8171	-0.077	65.6256	-11.7017		
25	4.3258	1.8867	-0.1355	77.0408	-24.4347		
30	4.476	1.9493	-0.1788	88.9726	-37.6666		
40	4.7471	2.0621	-0.2496	115.3785	-70.015		
50	5.	2.1674	-0.312	147.0423	-114.9488		
60	5.2529	2.2728	-0.3727	187.3957	-181.2693		
70	5.524	2.3856	-0.4366	243.0123	-287.5766		
75	5.6742	2.4482	-0.4717	280.6493	-368.3851		
80	5.8415	2.5178	-0.5106	329.4667	-482.9594		
84	5.9944	2.5815	-0.546	381.508	-616.3926		
90	6.2817	2.7011	-0.6123	502.4973	-967.5551		
95	6.6452	2.8525	-0.6957	712.0034	-1,694.882		
99	7.3268	3.1363	-0.8513	1,368.6063	-4,763.158		
Regression Statistics							
LC50	147.0423	LC50 Standard Error	221.9314				
LC50 LCL	13.9977	LC50 UCL	1,544.6442				
Log10[LC50]	2.1674	Standard Error	0.5211				
Beta	2.4016	Intercept	-0.2054				
Beta Standard Error	1.9318						

Probit Analysis - Least squares [Normal Distribution]				
Dose (Stimulus)	Actual Percent (%)	N	Probit (Y)	Weight (Z)
5.	0.025	10.	3.0396	1.0792
10.	0.025	10.	3.0396	1.0792
20.	0.025	10.	3.0396	1.0792
30.	0.1	10.	3.7183	2.6548
50.	0.1	10.	3.7183	2.6548
Regression Statistics				
LC50	119.6466	LC50 Standard Error		18.5731
LC50 LCL	80.7726	LC50 UCL		158.5206
Beta	0.017	Intercept		2.9629
Beta Standard Error	0.021			
LC10	44.3663	LC16	60.9132	
LC84	178.38	LC90	194.9269	
LC100	207.7467			

Table 8: 48 Hours Probit analysis.

Alpha value (for confidence interval)	0.05						
Probit Analysis - Finney Method [Lognormal Distribution]							
Log10[Dose (Stimulus)]	Actual Percent (%)	Probit Percent (%)	N	R	E(R)	Difference	Chi-square
0.699	0.025	0.0321	10	0.25	0.3213	-0.0713	0.0158
1.	0.1	0.0975	10	1.	0.9745	0.0255	0.0007
1.301	0.3	0.229	10	3.	2.2904	0.7096	0.2199
1.4771	0.4	0.338	10	4.	3.3803	0.6197	0.1136
1.699	0.4	0.4962	10	4.	4.9624	-0.9624	0.1866
Chi-square							
Chi-square	0.5366						
Degrees Of Freedom	3						
p-level	0.9108						
Dose (Stimulus) Percentile							
Percentile	Probit (Y)	Log10 [Dose (Stimulus)]	Standard Error	Dose (Stimulus)	Standard Error		
1	2.6732	0.4402	0.846	2.7554	9.467		
5	3.3548	0.8104	0.5278	6.4628	9.9367		
10	3.7183	1.0079	0.3665	10.1826	9.6507		
16	4.0056	1.1639	0.253	14.5856	8.9852		
20	4.1585	1.247	0.2057	17.661	8.6796		
25	4.3258	1.3379	0.1763	21.7707	9.0829		
30	4.476	1.4195	0.1793	26.2697	11.1588		
40	4.7471	1.5667	0.2466	36.8721	22.0777		
50	5.	1.7041	0.3442	50.593	44.4362		
60	5.2529	1.8415	0.454	69.4197	86.5293		
70	5.524	1.9887	0.5773	97.4375	171.1597		
75	5.6742	2.0703	0.6469	117.5729	247.443		
80	5.8415	2.1612	0.7251	144.9327	371.1382		
84	5.9944	2.2443	0.7971	175.4917	535.9255		
90	6.2817	2.4003	0.9331	251.3742	1,062.8553		
95	6.6452	2.5978	1.1063	396.062	2,513.8633		
99	7.3268	2.968	1.4327	928.9484	12,560.913		
Regression Statistics							
LC50	50.593	LC50 Standard Error		44.4362			
LC50 LCL	29.3901	LC50 UCL		657.1199			
Log10[LC50]	1.7041	Standard Error		0.3442			
Beta	1.841	Intercept		1.8628			
Beta Standard Error	0.7079						

Probit Analysis - Least squares [Normal Distribution]				
Dose (Stimulus)	Actual Percent (%)	N	Probit (Y)	Weight (Z)
5.	0.025	10.	3.0396	1.0792
10.	0.025	10.	3.0396	1.0792
20.	0.025	10.	3.0396	1.0792
30.	0.1	10.	3.7183	2.6548
50.	0.1	10.	3.7183	2.6548
Regression Statistics				
LC50	119.6466	LC50 Standard Error	18.5731	
<i>LC50 LCL</i>	80.7726	<i>LC50 UCL</i>	158.5206	
<i>Beta</i>	0.017	<i>Intercept</i>	2.9629	
<i>Beta Standard Error</i>	0.021			
<i>LC10</i>	44.3663	<i>LC16</i>	60.9132	
<i>LC84</i>	178.38	<i>LC90</i>	194.9269	
<i>LC100</i>	207.7467			

Table 9: 72 Hours Probit analysis.

<i>Alpha value (for confidence interval)</i>	0.05						
Probit Analysis - Finney Method [Lognormal Distribution]							
Log10[Dose (Stimulus)]	Actual Percent (%)	Probit Percent(%)	N	R	E(R)	Difference	Chi-square
0.699	0.025	0.0457	10	0.25	0.4572	-0.2072	0.0939
1.	0.2	0.1698	10	2.	1.6978	0.3022	0.0538
1.301	0.5	0.4121	10	5.	4.1206	0.8794	0.1877
1.4771	0.6	0.5818	10	6.	5.8177	0.1823	0.0057
1.699	0.7	0.7723	10	7.	7.7231	-0.7231	0.0677
Chi-square							
<i>Chi-square</i>	0.4088						
<i>Degrees Of Freedom</i>	3						
<i>p-level</i>	0.9384						
Dose (Stimulus) Percentile							
Percentile	Probit (Y)	Log10[Dose (Stimulus)]	Standard Error	Dose (Stimulus)	Standard Error		
1	2.6732	0.4365	0.3647	2.7321	2.5733		
5	3.3548	0.7165	0.2598	5.2058	3.3028		
10	3.7183	0.8658	0.2064	7.3418	3.6217		
16	4.0056	0.9838	0.1669	9.6344	3.7953		
20	4.1585	1.0467	0.1477	11.1343	3.8598		
25	4.3258	1.1154	0.1289	13.0429	3.9269		
30	4.476	1.1771	0.1148	15.0339	4.0221		
40	4.7471	1.2884	0.1002	19.4278	4.5237		
50	5.	1.3923	0.1031	24.679	5.911		
60	5.2529	1.4962	0.1207	31.3495	8.8278		
70	5.524	1.6076	0.1502	40.512	14.2968		
75	5.6742	1.6693	0.1693	46.6961	18.6737		
80	5.8415	1.738	0.192	54.7007	24.9853		
84	5.9944	1.8008	0.2137	63.2164	32.3819		
90	6.2817	1.9189	0.256	82.9571	51.7879		
95	6.6452	2.0682	0.3114	116.9958	91.2509		
99	7.3268	2.3482	0.4179	222.9238	249.1922		
Regression Statistics							
LC50	24.679	LC50 Standard Error	5.911				
<i>LC50 LCL</i>	16.5872	<i>LC50 UCL</i>	42.0465				
<i>Log10[LC50]</i>	1.3923	<i>Standard Error</i>	0.1031				
<i>Beta</i>	2.4343	<i>Intercept</i>	1.6106				
<i>Beta Standard Error</i>	0.6772						

Probit Analysis - Least squares [Normal Distribution]				
Dose (Stimulus)	Actual Percent (%)	N	Probit (Y)	Weight (Z)
5.	0.025	10.	3.0396	1.0792
10.	0.2	10.	4.1585	3.8171
20.	0.5	10.	5.	5.
30.	0.6	10.	5.2529	4.7471
50.	0.7	10.	5.524	4.452
Regression Statistics				
LC50	29.1742	LC50 Standard Error	6.0467	
<i>LC50 LCL</i>	16.9437	<i>LC50 UCL</i>	41.4048	
<i>Beta</i>	0.037	<i>Intercept</i>	3.9211	
<i>Beta Standard Error</i>	0.0153			
<i>LC10</i>	-5.4857	<i>LC16</i>	2.1327	
<i>LC84</i>	56.2158	<i>LC90</i>	63.8341	
<i>LC100</i>	69.7365			

Table 10: 96 Hours Probit analysis.

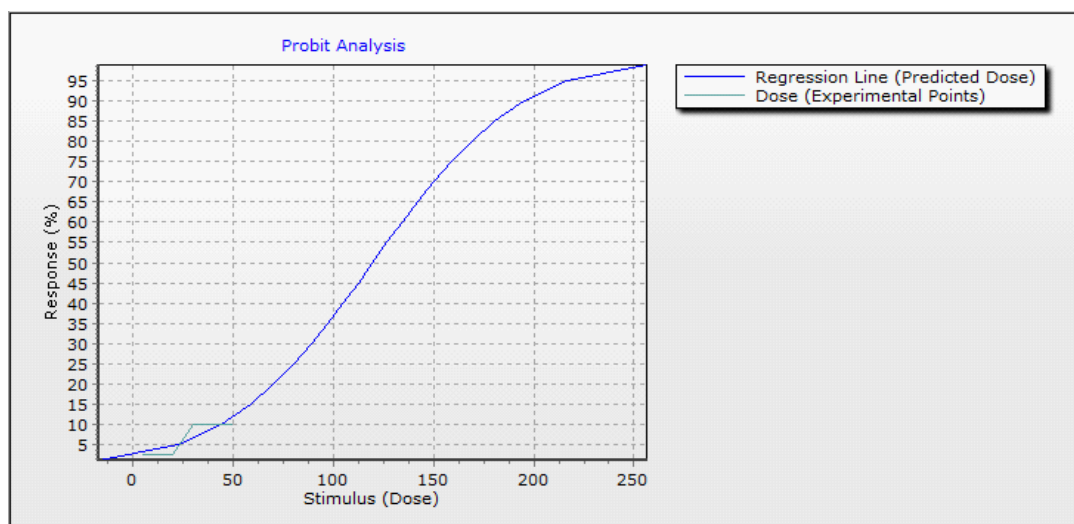


Figure 2: 48 Hours Probit analysis.

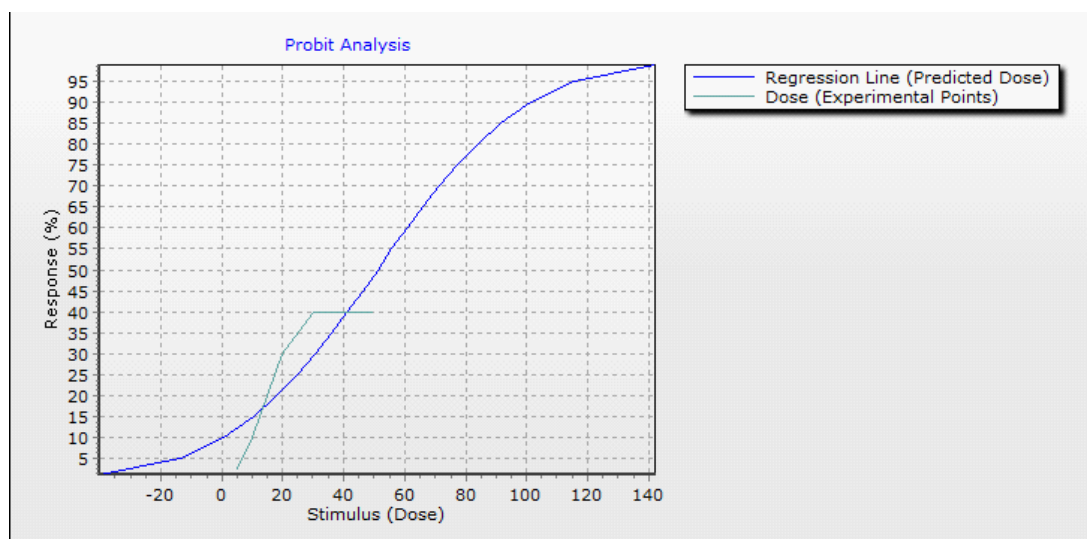


Figure 3: 72 Hours Probit analysis.

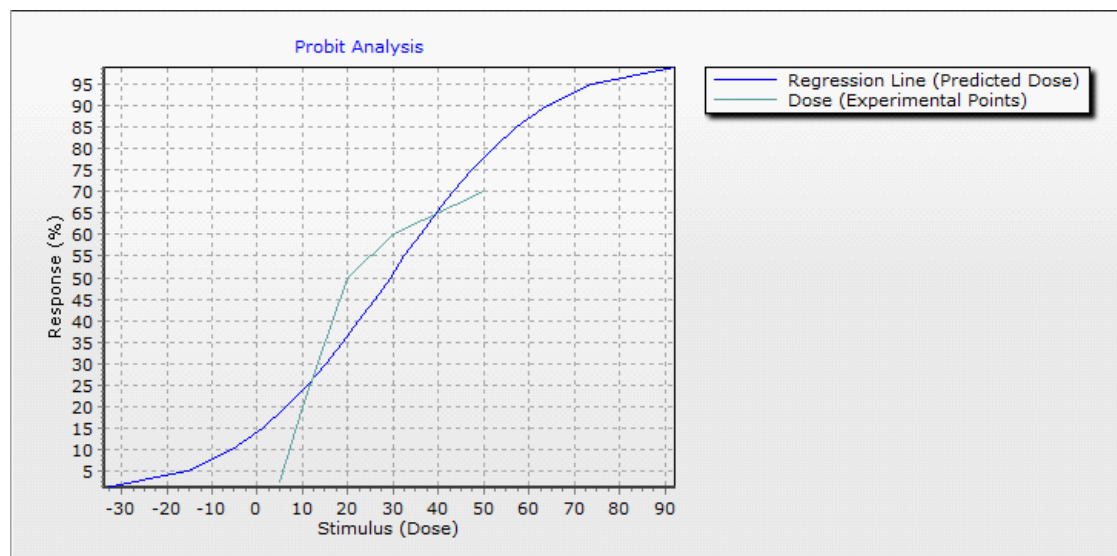


Figure 4: 96 Hours probit analysis.

mixture with a dilution factor of 20, whereas, zero mortality (0%) was observed in effluent mixture with a dilution factor of 200 at 96 hr.

Earlier studies by Sridevi and Kanmani [6], by taking effluents from other textile dyeing industries in Erode area with a dilution factor of 5 showed high mortality of 100% and zero mortality (0%) at dilution factor of 20. Our results in the present study showed that the response of the penaeid prawns (*Metapanaeus monoceros*) to RO rejects from the Textile dye industrial effluents (TDIE) are more toxic.

The LC_{50} for effluent sample at 48 hr exposure period was 120 ppm, the upper limit is 159 ppm and the lower limit is 80 ppm. The LC_{50} for 72h exposure period was 51 ppm, the upper limit is 69 ppm and lower limit is 33 ppm. The LC_{50} for 96h exposure period was 29 ppm, the upper limit is 41 ppm and the lower limit is 17 ppm.

The study shows that the overall toxicity increases when a number of effluents are mixed together. It is therefore concluded that the effluent mixture from 736 textile dyeing industries in Erode should be thoroughly treated before it is discharged through a common effluent pipe line.

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