Synergistic effect of elevated temperature, pCO2 and nutrients on marine biofilm

Lalita V. Baragi, Arga Chandrashekar Anil*

Academy of Scientific and Innovative Research (AcSIR) CSIR- National Institute of Oceanography, Dona Paula, Goa 403 004, India

*Correspondence to: Dr. A. C. Anil, Tel.: +91(0)832-2450404; fax: +91(0)832-2450615, Email:acanil@nio.org

Abstract

Natural marine biofilms provide signatures of the events that occur over a period of time and can be used as bioindicators of environmental changes. Hence, the effects of temperature (30 and 34°C), pCO_2 (400 and 1500 µatm) and nutrients (unenriched and enriched f/2 media) on the marine biofilm were evaluated using a 2×2×2 factorial design. In unenriched condition, acidification significantly increased the abundance of phytoperiphytes whereas reduced that of bacteria and it was vice versa in the enriched condition. Warming had significant negative effect on the abundance of both phytoperiphytes and bacteria, except in unenriched condition wherein it favoured bacterial growth. Synergistically, acidification and warming had deleterious effects resulting in further reduction in the abundance of both phytoperiphytes and bacteria, except in enriched condition wherein bacterial abundance increased. Such changes in biofilm communities in response to warming and acidification can have cascading effect on the subsequent build-up of macrofouling community.

Keywords: Biofilm, warming, acidification, nutrient enrichment, phytoperiphyton, bacteria

Introduction

Diatoms, cyanobacteria (autotrophic) and heterotrophic bacteria are the major periphytic communities in a biofilm (Costerton et al., 1995), whose complex biological systems play a major role in benthic trophodynamics (Sommer, 1999) and also provide cues for the settlement of higher organisms (Qian et al., 2007). Biofilm formation, growth and development are dependent on temperature, pCO₂ and nutrient availability in the surrounding waters (Melo and Bott, 1997; Qian et al., 2007; Russell et al., 2013). Biofilm communities reorganize and respond rapidly to the environmental changes (Tolker-Nielsen and Molin, 2000), thus providing signatures of the events that occur over a period of time. As such biofilms can be used as bioindicators of environmental changes (Burns and Ryder, 2001 Sabater and Admiraal, 2005).

Warming and acidification are considered as major threats to the marine biota (Stocker et al., 2013). These processes are the consequences of the CO_2 emission by anthropogenic activities. While the impact of warming on marine biota is well known, the acidification might bring more complicated irreparable changes in the ecosystem. Climatologists have predicted a further three times increase in p CO_2 concentration, thus resulting in pH drop by 0.3 - 0.4 units and temperature increase by 1-4 °C by the end of this century (Stocker et al., 2013). Studies conducted so far reported varying effects of these stressors on microalgae and bacteria, however the results are either conflicting or scarce. Therefore, it is very difficult to draw a conclusion or to predict the patterns of future ecological impacts. Moreover, the potential effects of acidification on biofilms are still completely unknown (Cartaxana et al., 2015; Johnson et al., 2013; Lidbury et al., 2012; Russell et al., 2013).

So far, studies done on the effects of warming and acidification on biofilms are skewed towards unialgal diatoms (Torstensson et al., 2012; Torstensson et al., 2013); but, very few snapshots are available on multispecies natural biofilms (Cartaxana et al., 2015; Lidbury et al., 2012; Russell et al., 2013). In an earlier investigation using a unialgal fouling diatom (*Navicula distans*) it was observed that warming and acidification synergistically had a significant negative impact on its growth (Baragi et al., 2015). In the case of natural biofilms, complex interaction exists between different types of periphytic organisms and thus, it is hypothesised that the discrete response of each component to these environmental changes could modify the behavioural response of the whole biofilm.

Nutrients and CO_2 play vital role in the growth of photoautotrophs in a biofilm matrix. It has been reported that their influence on the photoautotrophic community can be better understood by using a nutrient-load hypothesis (Brauer et al., 2012), modified by Verspagen et al. (2014). The modified nutrient-load hypothesis describes the resource limitation patterns as a function of two limiting factors: nutrients (nitrate, phosphate and silicate) and carbon. It predicts that at low nutrient concentration and high pCO₂, nutrient will be the limiting resource for microalgae; whereas, at high nutrient concentration and low pCO₂, carbon will become limiting. However, both high nutrients and high pCO₂ will produce dense microalgal population that will increase the turbidity of water column thus creating a light limited condition.

The present study was designed to assess the response of natural biofilm community to temperature, pCO_2 and nutrients at community level. Biofilms were grown for 2 days under natural conditions (in the field) and their subsequent development in the laboratory under different combinations of

temperature, pCO_2 and nutrients to address the following questions (a) How does acidification and warming affect the natural biofilm communities (diatom, picopheriphytes and bacteria)? and (b) How does the nutrient concentration affect the response of biofilm community to acidification and warming?

2. Materials and Methods

2.1. Natural biofilm development

Natural biofilms were developed (April, 2013) by submerging polystyrene slides $(7.6 \times 2.5 \text{ cm})$ for 2 days at Dona Paula $(15^{\circ}.31^{\circ}N, 73^{\circ}.59^{\circ}E)$, west coast of India. For this purpose, a total of 45 slides were fitted in a slide holder and suspended in the subtidal environment (depth ~ 1 m). After 2 days, the slide holders were retrieved and slides were used for the laboratory experiment. Out of 45 slides, 40 slides were used for the experiment [5 slides (5 replicates) X 8 treatments] and 5 slides for day 0 analysis.

2.2. Manipulation of pCO_2 and temperature levels

The two pCO₂ levels used in the present study were ~400 µatm (present day pCO₂) and ~1500 µatm (pCO₂ level projected for the year 2100) (Stocker et al., 2013). The projected 2100 year pCO₂ concentration was manipulated by adding NaHCO₃ and HCl (0.1 N) following the method described by Riebesell et al. (2010) in which total alkalinity (TA) was kept constant, but dissolved inorganic carbon (DIC) was increased. This is one of the methods recommended for manipulating the seawater CO₂ for small-volume experiments.

TA and pH values of the seawater were used to determine the pCO₂ concentration. TA was quantified by Gran titration (Gran, 1952) and the seawater pH was measured using a pH electrode calibrated using buffers (National Bureau of Standards). The TA, pH, temperature and salinity measurements were used to calculate the carbonate chemistry parameters of the treatments using CO₂ Sys program developed by Lewis and Wallace (1998), using the dissociation constant of Mehrbach et al. (1973) and then modified by Dickson and Millero (1987).

The two temperature levels used in the present study were $\sim 30^{\circ}$ C (ambient - present day temperature) and $\sim 34^{\circ}$ C µatm (elevated - projected for the year 2100). Since the average annual temperature of Dona Paula Bay (Goa, India), where the biofilm was developed, is 30°C, it was used as present day temperature. Incubators (REMI, Microprocessor cooling incubator) were used to maintain the experimental temperatures, with $\pm 0.05^{\circ}$ C accuracy.

2.3. Experimental design in the laboratory

The experiment was conducted to test the effects of temperature (30°C and 34°C), pCO₂ (400 µatm and 1500 µatm) and nutrients (unenriched and enriched f/2 media) on the response of marine biofilms, using a 2 × 2 × 2 factorial design (total 8 combinations) on the marine biofilm. f/2 media was used for enriched conditions. The condition with ambient temperature (30°C) and pCO₂ (400 µatm) with unenriched nutrients was considered as control condition. The average initial concentrations of NO₃+NO₂, PO₄ and Si(OH)₄ in unenriched seawater were 3.0 ± 0.4 , 2.0 ± 0.5 , and 13.1 ± 1.3 µmol l⁻¹ respectively; whereas, in enriched seawater were 238 ± 43 , 45 ± 11 and 101.3 ± 32 μ mol l⁻¹, respectively. In the laboratory, the 2 days old natural biofilms were incubated in the 500 ml polycarbonate bottles, at different treatment condition. The polycarbonate bottles were incubated under 70 mmol photons m⁻² s⁻¹ illumination with a 12Light: 12Dark photoperiod, for 10 days. A total of 45 slides were used in the experiment (5 slides X 8 treatments = 40 slides and 5 slides for day 0 analysis). Every day the slides were transferred to new 500 ml polycarbonate bottles containing seawater of appropriate condition, this was done to maintain the pCO₂ and nutrient levels stable. Five slides from each treatment were sampled, on day 10 of the experimental period, and used for analysis.

2.4. Biofilm abundance and community analysis

In this paper the terms 'periphytes' refer to both autotrophic and heterotrophic organisms in the biofilm, 'picoperiphytes' to both autotrophic and heterotrophic organisms with cell size $\leq 2\mu m$, 'phytoperiphytes' to autotrophic organisms in the biofilm, and 'picophytoperiphytes' to autotrophic organisms with cell size $\leq 2\mu m$.

Diatoms were enumerated, by direct light microscopy of 400X magnification, in 10 randomly selected fields (area = 0.38 dm^2) on each slide used for natural biofilm development. Diatoms were identified to genus level based on standard identification keys (Tomas, 1996). Final counts are expressed as diatoms dm⁻². This data are used to determine the Margalef's generic richness (d), Pielou's evenness (J') and Shannon-Wiener's diversity (H') of diatoms.

After diatom analysis each slide was scraped, with 10 ml of 0.22 μ m autoclaved filtered sea water, using a nylon brush as this method is identified to be efficient in removing the adhered cells (Patil and Anil, 2005); subsequently, the scrapped material was sonicated. For picophytoperiphyton and heterotrophic bacterial analysis, the sonicated sample (1 ml for each) was preserved with paraformaldehyde (0.2% final concentration) and stored at -80 °C until further analysis. The frozen samples were thawed and analysed through a BD FacsAria II flowcytometer equipped with a laser emitting at 488 nm following standard methods (Marie et al., 1997; Mitbavkar et al., 2012). For heterotrophic bacteria, samples were analysed after staining with nucleic acid stain, SYBR Green I (final concentration 10^{-4}) (Marie et al., 1997).

2.5. Chlorophyll a (Chl a) and photosynthetic quantum yield analysis

Chl *a* estimation was done by filtering a known volume of sample (5 ml) onto the GF/F filter paper and extracting the pigments in 90% acetone for 24 hrs at 4 °C. Absorbance of the extract was measured in a spectrophotometer. These values were then used to calculate the chl *a* concentration using the method described by Parsons et al. (1984).

The status of chlorophyll in the biofilms under different treatments was assessed using Fluorescence Induction and Relaxation (FIRe) technique. The FIRe fluorometer measures the initial (F0), maximum (Fm) and variable (Fv = Fm-F0) chlorophyll fluorescence, which is later used to measure photosynthetic quantum yield (Fv/Fm). Fv/Fm can be used to assess the physiological status of the photosynthesizing organisms in the biofilm. Lower values of Fv/Fm indicates unhealthy cells and vice versa (Patil and Jagadeesan, 2011).

2.6. Carbon contribution of different biofilm communities

Autotrophic carbon biomass was derived from chl *a* using a carbon to chlorophyll ratio of 40 (Gallegos, 2001). This value was used as the biofilm was dominated by diatoms. Picophytoperiphyton carbon biomass was calculated by measuring the carbon biomass from each of the 3 picophytoperiphyton groups. For measuring the carbon biomass from each group, initially right angle light scatter (RALS) data obtained from flowcytometry was converted to cell diameter (d) as explained by Worden et al. (2004). Later, the cell diameter was used to estimate the biovolume using the equation, $V=\pi/6$ (d³), assuming that picophytoperiphyton cells are spherical. The factor to convert cell abundance to carbon biomass for different picophytoperiphyton groups was derived from the biovolume to carbon conversion factor of 254 fg C mm³ (Baudoux et al., 2007). From these estimations, the percentage contribution of picophytoperiphyton (< 2 µm) and larger phytoperiphyton (> 2 µm) to total carbon biomass was calculated. Heterotrophic bacterial carbon biomass was calculated using a carbon conversion factor of 20 fg C cell⁻¹ (Lee and Fuhrman, 1987).

2.7. Data analyses

Margalef's species richness (d), Pielou's evenness (J') and Shannon-Wiener's diversity (H') of the fouling diatom community were calculated, using PRIMER software version 6. Prior to statistical analysis, data was checked for normality by Shapiro-Wilk test and homogeneity of variances by Levens test to meet the assumptions of parametric tests. Data for each taxonomic group (diatom, picophytoperiphyton and bacteria) were log-transformed to meet assumptions of parametric tests. The variation in the descriptors of biofilm (abundance of diatom, picophytoperiphyton, and bacteria, chl *a* concentration and Fv/Fm ratio) was analysed by three-way MANOVA, with temperature (30 and 34° C), pCO₂ (400 vs 1500 µatm) and nutrients (Enriched and unenriched) as factors. Due to a significant 3-way interaction, we performed 2-way ANOVAs on each nutrient level (unenriched and enriched) independently, to investigate the effect of pCO₂ (400 vs 1500 µatm) and temperature (30 and 34° C) on the responses of biofilm. All statistical analysis was performed using STATISTICA software.

3. Results

At ambient temperature (30 °C), the initial ambient and elevated pCO₂ was 514±19 µatm and 1665±18 µatm, respectively. At elevated temperature (34 °C), the initial ambient and elevated pCO₂ was 493±29 µatm and 1560±36 µatm, respectively. The temperature and pCO₂ at ambient and elevated levels were significantly different, throughout the experiment (p < 0.05). The seawater pCO₂ concentration and nutrients were kept stable by every day transferring the slides to new polycarbonate bottles containing seawater of appropriate conditions. The three-way MANOVA revealed significant main effects of temperature, pCO₂ and nutrients; and interactions between these factors on the biofilm (Wilks' Lambda: P < 0.001).

3.1. Chlorophyll a and photosynthetic quantum yield (Fv/Fm)

The Chl *a* concentration on day 0 was $0.53 \pm 0.12 \mu \text{g dm}^{-2}$ (Fig. 1). On day 10, the maximum Chl *a* concentration (99.74 ± 4.36 $\mu \text{g dm}^{-2}$) was observed at 514 μatm and 30°C condition with enriched nutrients (ANOVA, Post Hoc Tukey's HSD p < 0.001). The rise in pCO₂ to 1665 μatm at 30°C

increased the chl *a* concentration in both unenriched $(3.03 \pm 0.47 \ \mu g \ dm^{-2})$ and enriched conditions $(16.92 \pm 0.62 \ \mu g \ dm^{-2})$; however the pCO₂ effect was significant only in enriched condition (p < 0.001). Temperature had significant negative effect on chl *a* concentration (p < 0.001) with lower concentration at elevated temperature (34°C) (p < 0.001); however, compared to unenriched condition the concentration was higher in enriched condition (ANOVA, Tukey's HSD p < 0.001).

On day 0, the mean Fv/Fm was 0.75 (Fig. 2). On day 10, Fv/Fm values of biofilms at 514 μ atm and 30°C were not significantly different in both enriched and unenriched conditions (p = 0.872). Irrespective of nutrient levels, the value was lower at 1665 μ atm than 514 μ atm pCO₂ (p < 0.01). Significant negative effect of temperature was observed only in unenriched condition, with lower values in treatments with 34°C (p < 0.001).

3.2. Periphytic community

3.2.1. Autotrophs

3.2.1.1. Diatoms

The mean diatom abundance on day 0 was 1.9×10^4 cells dm⁻² (Fig. 3A). On day 10, the abundance trend was similar to chl *a* concentration with highest abundance at 514 µatm pCO₂ and 30°C in enriched condition (10×10^5 cells dm⁻²) (ANOVA, Post Hoc Tukey's HSD p < 0.001). Irrespective of nutrient levels, the higher pCO₂ (1665 µatm) at 30°C enhanced the diatom abundance: unenriched (1.7×10^5 cells dm⁻²) and enriched (3.4×10^5 cells dm⁻²). However, in enriched condition the abundance was lower (~77%) compared to 514 µatm pCO₂ and 30°C (p < 0.001). Temperature significantly reduced the abundance of diatoms in conditions with 34°C and with a combination of 1560 µatm and 34°C compared to 514 µatm pCO₂ and 30°C (p < 0.01). In the synergistic condition of elevated pCO₂ (1560 µatm) and temperature (34°C), the diatom abundance was 2-3 fold higher in enriched than unenriched conditions.

The diatom community composition was altered by temperature and pCO₂ conditions (Fig. 4A). Diatom belonging to 12 genera were recorded with *Navicula* as the dominant genus in all the treatments, throughout the experiment. At 30°C, either enrichment of nutrients or pCO₂ increased the diatom abundance whereas nutrients and pCO₂ synergistically decreased the diatom abundance except *Thalassionema* spp and *Achnanthes* spp. Irrespective of nutrient concentrations, increased temperature either reduced the diatom abundance (*Navicula* spp., *Nitzchia* spp., *Amphora* spp. and *Pleurosigma* spp.) or eliminated the diatoms (*Striatella* spp., *Thalassionema* spp., *Licmophora* spp. *Melosira* spp.) in the biofilm.

Changes in the diatom community were recorded in terms of Shannon-Wiener diversity (H'), Margalef's species richness (d), and Pielou's evenness (J'). In both the nutrient levels, the mean generic diversity (H') of fouling diatom significantly varied with temperature with lower diversity at 34°C (p < 0.001). The diversity was similar in both the pCO₂ levels (p > 0.05). Biofilm grown at 514 µatm and 30°C had highest diversity (0.62 ± 0.01) compared to other treatments (ANOVA, Tukey's HSD, p < 0.05). The influence of temperature and pCO₂ on diatom generic evenness (J') was significant only in enriched condition with highest (0.79 ± 0.1) in condition with 1560 µatm and 34°C (p < 0.05). Diatom generic richness (J') was significantly influenced by temperature and pCO₂ in both

the nutrient conditions with a lower richness in conditions with 34°C and combination of 1560 μ atm pCO₂ and 34°C (p < 0.001).

3.2.1.2. Picophytoperiphytes

The mean picophytoperiphytic abundance on day 0, was 2×10^4 cells dm⁻² (Fig. 3B). Picophytoperiphytic community comprised of Synechococcus spp., Prochlorococcus-like cells and picoeukaryotes. Irrespective of treatments, Synechococcus spp. was the dominant group (Fig. 4B). On day 10, the total picophytoperiphytic abundance was highest in enriched condition at 514 µatm and 30° C (11 × 10⁶ cells dm⁻²) (p < 0.001) with *Prochlorococcus*-like cells as dominant group followed by picoeukaryotes. Regardless of nutrient levels at 30°C, elevated pCO₂ (1665 µatm) enhanced the picophytoperiphytic abundance: unenriched (5.5 $\times 10^5$ cells dm⁻²) and enriched (3 $\times 10^6$ cells dm⁻²). However, in enriched condition the abundance was lower (\sim 73%) compared to 514 µatm pCO₂ (p < 0.001). In unenriched condition, all the groups showed increasing trend at elevated pCO_2 (1665 µatm) compared to 514 µatm; Synechococcus spp. (~58%), Prochlorococcus-like cells (~35%) and picoeukaryotes (~145%); however in enriched condition only Synechococcus spp. increased (10 fold). The total picophytoperiphytic abundance also increased in conditions with 34°C and with a combination of 1560 µatm and 34°C compared to 514 µatm pCO₂ and 30°C (p < 0.01). The total picophytoperiphytic abundance in enriched conditions was 30-60% higher than unenriched conditions, however was 94% lower than condition with ambient pCO₂ and temperature (514 µatm pCO₂ and 30°C). The abundance of only Synechococcus spp. increased in these conditions irrespective of nutrient conditions.

3.2.2. Heterotroph

3.2.2.1. Bacteria

The mean initial bacterial abundance was 1.7×10^7 cells dm⁻² (Fig. 3C). On day 10, the enrichment of nutrients at 514 µatm pCO₂ and 30°C increased the abundance by ~23% compared to unenriched condition (p < 0.001). Irrespective of temperature treatments, elevated pCO₂ decreased the bacterial abundance in unenriched condition whereas increased the abundance in enriched condition (p < 0.01). Temperature had significant effect on bacterial abundance with increasing and decreasing bacterial abundance with temperature in unenriched and enriched conditions, respectively (Fig. 3C, p < 0.001).

3.3. Diatom, picophytoperiphytic and bacterial carbon contribution

The variation in phytoperiphytic (autotrophic) and bacterial carbon biomass is presented in Fig. 5A and B. Temperature and nutrients significantly influenced the pico (< 2 μ m) and larger phytoperiphytic (> 2 μ m) carbon contribution to total autotrophic carbon biomass (Fig. 6). Irrespective of pCO₂ levels, the picophytoperiphytic carbon contribution was lower in conditions with 30°C: unenriched (0.07-0.1 %) and enriched (0.06-0.1%); whereas the contribution increased at elevated temperature (34°C): unenriched (1.07 %) and enriched (0.27 %) and with combination of 1560 μ atm pCO₂ and 34°C: unenriched (2.03 %) and enriched (0.26 %).

Bacterial biomass (BB) to autotrophic carbon biomass (AB) ratio (BB:AB) was significantly influenced by temperature and nutrients (p < 0.05), with a higher ratio at elevated temperature (34°C): unenriched (0.078-0.102) and enriched (0.004-0.011) (Fig. 5C).

4. Discussion

The present study highlights that acidification, warming and elevated nutrient significantly affect the biofilm community. The altered community structure in response to these factors is indeed a function of either the direct effect of these factors on the physiology of individual organisms or the effect of species interaction, such as competition, among the different groups within the biofilm under different experimental treatments.

Nutrient enrichment increased the abundance of phytoperiphytes (diatom and picophytoperiphyte) and bacteria in the biofilm, which corroborates the earlier studies (Cermeño et al., 2008; Chiu et al., 2008; Falkowski et al., 1998; Kirchman, 1994). According to nutrient-load hypothesis (Verspagen et al., 2014), at lower pCO₂ and nutrient concentration, both carbon and nutrients act as limiting factors for microalgal growth, resulting in lower abundance as observed in the present study. However, at high nutrient and low pCO₂ condition, carbon behaves as the sole limiting factor for microalgae (Verspagen et al., 2014). In contrast to this, recent studies have observed that carbon is not a limiting factor for microalgal growth at ambient atmospheric CO₂ level (400 μ atm) because of the presence of the Carbon Concentrating Mechanisms (CCMs) which actively utilize bicarbonate and converts it to CO₂ (Giordano et al., 2005). This is also supported by the present study wherein, under ambient pCO₂ and enriched nutrients phytoperiphytes attained highest abundance.

Picophytoperiphytes dominated the phytoperiphytic community irrespective of nutrient levels, the reason of which could be attributed to their competitive strategies (Donald et al., 1997; Fogg, 1986). Because of the smaller size, picophytoperiphytes acquire and use nutrients more efficiently than larger phytoperiphytes (Donald et al., 1997; Fogg, 1986). Nutrient enrichment increased the abundance of picoperiphytes, however, its contribution to total carbon biomass was similar to unenriched condition. This could be due to the dominance of larger sized cells, *Synechococcus* spp. in unenriched condition whereas smaller sized *Prochlorococcus*-like cells in enriched condition. However, further studies should advocate this contrasting pattern of these two picophytoperiphytes in response to nutrient requirements.

In the present study, the abundance of phytoperiphytes increased in response to elevated pCO₂ which is in accordance with previous studies (Fu et al., 2007; Hicks et al., 2011; Johnson et al., 2013; Newbold et al., 2012). This is because, some microalgae take advantage of rising CO₂ level through downregulation of CCMs that cause reduction in the energetic cost and also by improving resource allocation (Hopkinson et al., 2011; Rost et al., 2008; Trimborn et al., 2009). At elevated pCO₂ and enriched condition phytoperiphytic (diatom and picophytoperiphyte) abundance increased supporting the nutrient-load hypothesis (Verspagen et al., 2014). However, the percentage increase was significantly lower than that at ambient pCO₂, the reason of which could be attributed to 3 possibilities: (1) the phytoperiphytes are light-limited as evidenced from the presence of stalked preiphytic diatom *Achnanthes* spp., and larger sized picophytoperiphyton such as *Synechococcus* spp. compared to smaller sized such as *Prochlorococcus*-like cells at ambient pCO₂; which might have resulted in overcrowding in the biofilm thus resulting in light-limiting conditions. (2) The lower pH, as a result of increased pCO₂, affects the acid base balance (Beardall et al., 2009) and enhance mitochondrial respiratory (Wu et al., 2010) or photorespiratory (Gao et al., 2012) processes in microalgae hence demanding high cellular energy which might have reduced the phytoperiphytic abundance at this condition. (3) The dominance of *Synechococcus* spp., which attains high growth rate at elevated pCO₂ compared to ambient pCO₂ (Fu et al., 2007) and is also known to produce substances with allelopathic effect on prokaryotic and eukaryotic cells (Martins et al., 2008) that may have inhibited the other periphytic organisms in the biofilm.

The addition of pCO_2 had no significant effect on the total bacterial abundance in enriched condition, which supports the earlier reports (Allgaier et al., 2008; Grossart et al., 2006; Oliver et al., 2014). However, under unenriched condition the lower abundance may be possibly due to the effect of lower pH which increases the respiration and in turn metabolic and energy cost of bacteria (Del Giorgio et al., 1999; Siu et al., 2014). This highlights the significant influence of nutrients on the response of bacteria to ocean acidification.

Elevated temperature, irrespective of pCO₂, had detrimental effect on biomass (Chl *a*), photosynthetic efficiency (Fv/Fm) and diatom abundance whereas the picophytoperiphytic abundance remained unaltered. Studies have reported negative effect of elevated temperature on microphytobenthic biomass (Cartaxana et al., 2015; Hicks et al., 2011). In the present study under warming condition, the magnitude of biomass reduction was higher in unenriched than enriched condition. Moreover, the Fv/Fm ratio, a proxy for phytoperiphytic health (Patil and Jagadeesan, 2011), also significantly decreased under unenriched condition. Thus, indicating that the magnitude of temperature impact on the phytoperiphytes depends on the nutrient availability.

It has been observed that the microalgae in the tropical environment are at their upper limit of temperature tolerance therefore, a slight increase can have detrimental effects on them (Thomas et al., 2012). Among diatoms, some were temperature tolerant with reduced abundance (*Navicula* spp., *Nitzchia* spp., *Amphora* spp. and *Pleurosigma* spp.) and others were temperature sensitive and were eliminated from the biofilm (*Striatella* spp., *Thalassionema* spp., *Licmophora* spp. *Melosira* spp). These differences in the response of diatoms to temperature may be attributed to species specific response to temperature (Eppley, 1972; Suzuki and Takahashi, 1995). It is also known that microalgae with a narrow range of temperature for growth are expected to be affected by the future rise in temperature (Beardall and Raven, 2004). Boyce et al. (2010) reported the rapid decline (~1% annaully) of global phytoplankton abundance with warming; with diatoms being the most affected group (Toseland et al., 2013).

Warming, under unenriched condition, favoured the growth of bacteria and this could be due to two reasons: (1) the positive effect of temperature on bacteria with higher bacterial production at elevated temperature (Apple et al., 2006), and (2) the availability of dissolved organic matter (DOM), source of nutrition for bacteria, from periphytic microalgae as microalgae are known to release high amounts of DOM under stressful conditions such as nutrient limitation. Under this condition, bacterial abundance was negatively correlated with phytoperiphytic abundance and had higher BB:AB ratio indicating less energy will be available to higher trophic organisms as the energy transfer from bacteria to higher trophic organisms is less efficient than phytoplankton to higher trophic organisms, because of the higher number of trophic levels in the food chain between bacteria and higher organisms.

In future, both factors, pCO₂ and temperature, will rise and make the ocean acidic as well as warmer. Under this condition with low nutrient concentration, diatoms attained lower abundance that may have resulted from combined stressful effects of warming and acidification. This could be due to higher energy requirement to cope up with an elevated temperature. Along with this, the additional energy is also required to meet the energy demand in maintaining the acid-base balance at lower pH. However, the addition of nutrients to this condition slightly increased the total phytoperiphytic abundance, thus, indicating that availability of nutrients compensated the negative effect of temperature and pCO₂.

Warming either alone and in synergy with acidification increased the percentage contribution of picophytoperiphytic carbon to total autotrophic carbon biomass under unenriched condition. This is in accordance with an earlier study (Agawin et al., 2000) wherein they reported that picophytoplankton dominate the biomass in the warm and nutrient poor waters. To sustain maximum growth rates, picophytoplankton require 10 fold less nutrient concentration compared to larger phytoplankton. Hence, under low nutrient concentration picophytoplankton are able to maintain high growth rate where larger phytoplankton are nutrient limited (Agawin et al., 2000). Additionally, the increase in pCO_2 had additive effect as it enhanced the growth of picophytoperiphytes, especially *Synechococcus* spp. (Fu et al., 2007). However, the lower contribution in enriched condition could be due to the increased numbers of larger phytoperiphytes.

This study utilizes a multispecies approach, natural periphytic organisms that naturally co-occur in the nature, which has advantages over single species to predict the response of the biofilm to anthropogenic drivers. This experiment however, does not fully reflect the natural environment as we have eliminated the impact of grazers. Nevertheless, the results allow us to understand the response of biofilm including the interaction among the components (autotrophs vs heterotrophs) occurring in the biofilm.

5. Conclusion

Our study points out a significant effect of warming on the biofilm growth and community composition than acidification, however the enrichment of nutrients decreased the magnitude of warming effect. Acidification and warming, individually and synergistically, resulted in altered biofilm community structure, which can have cascading effect on the subsequent build-up of macrofouling community. The response of microbial biofilms to the shifts in pCO_2 and temperature could be used as a biological indicator of future climatic scenario.

Acknowledgements

We are grateful to the Director of CSIR-National Institute of Oceanography, for his support and encouragement. We thank our colleagues for their help and suggestions. Lalita V. Baragi acknowledges the research fellowship provided by the Council of Scientific and Industrial Research (CSIR) (India) and support provided by Academy of Scientific and Innovative Research (AcSIR). This work was supported by project Ocean Finder PSC 0105. This is an NIO contribution no.#.

References

Agawin, N.S., Duarte, C.M., Agusti, S., 2000. Nutrient and temperature control of the contribution of picoplankton to phytoplankton biomass and production. Limnol. Oceanogr. 45, 591-600.

Allgaier, M., Riebesell, U., Vogt, M., Thyrhaug, R., Grossart, H.-P., 2008. Coupling of heterotrophic bacteria to phytoplankton bloom development at different pCO₂ levels: a mesocosm study. Biogeosciences 5, 1007-1022.

Apple, J.K., Del Giorgio, P., Kemp, W.M., 2006. Temperature regulation of bacterial production, respiration, and growth efficiency in a temperate salt-marsh estuary. Aquat. Microb. Ecol. 43, 243.

Baragi, L.V., Khandeparker, L., Anil, A.C., 2015. Influence of elevated temperature and pCO₂ on the marine periphytic diatom *Navicula distans* and its associated organisms in culture. Hydrobiologia. 762, 127-142.

Baudoux, A.-C., Veldhuis, M.J., Witte, H.J., Brussaard, C.P., 2007. Viruses as mortality agents of picophytoplankton in the deep chlorophyll maximum layer during IRONAGES III. Limnol. Oceanogr. 52, 2519-2529.

Beardall, J., Raven, J.A., 2004. The potential effects of global climate change on microalgal photosynthesis, growth and ecology. Phycologia 43, 26-40.

Beardall, J., Stojkovic, S., Larsen, S., 2009. Living in a high CO₂ world: impacts of global climate change on marine phytoplankton. Plant. Ecol. Divers. 2, 191-205.

Boyce, D.G., Lewis, M.R., Worm, B., 2010. Global phytoplankton decline over the past century. Nature 466, 591-596.

Brauer, V.S., Stomp, M., Huisman, J., 2012. The nutrient-load hypothesis: patterns of resource limitation and community structure driven by competition for nutrients and light. Am. Nat. 179, 721-740.

Burns, A., Ryder, D.S., 2001. Potential for biofilms as biological indicators in Australian riverine systems. Ecological Management & Restoration, 2, 53-64.

Cartaxana, P., Vieira, S., Ribeiro, L., Rocha, R.J., Cruz, S., Calado, R., da Silva, J.M., 2015. Effects of elevated temperature and CO₂ on intertidal microphytobenthos. BMC Ecol. 15, 10.

Cermeño, P., Dutkiewicz, S., Harris, R.P., Follows, M., Schofield, O., Falkowski, P.G., 2008. The role of nutricline depth in regulating the ocean carbon cycle. Proc. Natl. Acad. Sci. 105, 20344-20349.

Chiu, J., Zhang, R., Wang, H., Thiyagarajan, V., Qian, P., 2008. Nutrient effects on intertidal community: from bacteria to invertebrates. Mar. Ecol. Prog. Ser. 358, 41.

Costerton, J.W., Lewandowski, Z., Caldwell, D.E., Korber, D.R., Lappin-Scott, H.M., 1995. Microbial biofilms. Annu. Rev. Microbiol. 49, 711-745.

Del Giorgio, P.A., Cole, J.J., Caraco, N.F., Peters, R.H., 1999. Linking planktonic biomass and metabolism to net gas fluxes in northern temperate lakes. Ecology 80, 1422-1431.

Dickson, A., Millero, F., 1987. A comparison of the equilibrium constants for the dissociation of carbonic acid in seawater media. Deep. Sea Res. 34, 1733-1743.

Donald, K.M., Scanlan, D.J., Carr, N.G., Mann, N.H., Joint, I., 1997. Comparative phosphorus nutrition of the marine cyanobacterium *Synechococcus* WH7803 and the marine diatom *Thalassiosira weissflogii*. J. Plankton Res. 19, 1793-1813.

Eppley, R.W., 1972. Temperature and phytoplankton growth in the sea. Fish Bull. 70, 1063-1085.

Falkowski, P.G., Barber, R.T., Smetacek, V., 1998. Biogeochemical controls and feedbacks on ocean primary production. Science 281, 200-206.

Fogg, G., 1986. Picoplankton. Proc. R. Soc. Lond. [Biol.] 228, 1-30.

Fu, F.X., Warner, M.E., Zhang, Y., Feng, Y., Hutchins, D.A., 2007. Effects of increased temperature and CO₂ on photosynthesis, growth, and elemental ratios in marine *Synechococcus* and *Prochlorococcus* (Cyanobacteria). J. Phycol. 43, 485-496.

Gallegos, C.L., 2001. Calculating optical water quality targets to restore and protect submersed aquatic vegetation: overcoming problems in partitioning the diffuse attenuation coefficient for photosynthetically active radiation. Estuaries 24, 381-397.

Gao, K., Xu, J., Gao, G., Li, Y., Hutchins, D.A., Huang, B., Wang, L., Zheng, Y., Jin, P., Cai, X., 2012. Rising CO₂ and increased light exposure synergistically reduce marine primary productivity. Nat. Clim. Change 2, 519-523.

Giordano, M., Beardall, J., Raven, J.A., 2005. CO₂ concentrating mechanisms in algae: mechanisms, environmental modulation, and evolution. Annu. Rev. Plant. Biol. 56, 99-131.

Gran, G., 1952. Determination of the equivalence point in potentiometric titrations. Part II. Analyst 77, 661-671.

Grossart, H.-P., Allgaier, M., Passow, U., Riebesell, U., 2006. Testing the effect of CO₂ concentration on the dynamics of marine heterotrophic bacteriplankton. Limnol. Oceanogr. 51, 1-11.

Hicks, N., Bulling, M.T., Solan, M., Raffaelli, D., White, P.C., Paterson, D.M., 2011. Impact of biodiversity-climate futures on primary production and metabolism in a model benthic estuarine system. BMC Ecol. 11, 7.

Hopkinson, B.M., Dupont, C.L., Allen, A.E., Morel, F.M., 2011. Efficiency of the CO₂-concentrating mechanism of diatoms. Proc. Natl. Acad. Sci. 108, 3830-3837.

Johnson, V., Brownlee, C., Rickaby, R., Graziano, M., Milazzo, M., Hall-Spencer, J., 2013. Responses of marine benthic microalgae to elevated CO₂. Mar. Biol. 160, 1813-1824.

Kirchman, D., 1994. The uptake of inorganic nutrients by heterotrophic bacteria. Microb. Ecol. 28, 255-271.

Lee, S., Fuhrman, J.A., 1987. Relationships between biovolume and biomass of naturally derived marine bacterioplankton. Appl. Environ. Microbiol. 53, 1298-1303.

Lewis, E., Wallace, D., 1998. Program Developed for CO₂ System Calculations (Carbon Dioxide Information Analysis Center, Oak Ridge National Laboratory, US Dept. of Energy, Oak Ridge, TN).

Lidbury, I., Johnson, V., Hall-Spencer, J.M., Munn, C.B., Cunliffe, M., 2012. Community-level response of coastal microbial biofilms to ocean acidification in a natural carbon dioxide vent ecosystem. Mar. Pollut. Bull. 64, 1063-1066.

Marie, D., Partensky, F., Jacquet, S., Vaulot, D., 1997. Enumeration and cell cycle analysis of natural populations of marine picoplankton by flow cytometry using the nucleic acid stain SYBR Green I. Appl. Environ. Microbiol. 63, 186-193.

Martins, R.F., Ramos, M.F., Herfindal, L., Sousa, J.A., Skærven, K., Vasconcelos, V.M., 2008. Antimicrobial and cytotoxic assessment of marine cyanobacteria-Synechocystis and *Synechococcus*. Mar. Drugs 6, 1-11.

Mehrbach, C., Culberson, C.H., Hawley, J.E., Pytkowicz, R.M., 1973. Measurement of the Apparent Dissociation Constants of Carbonic Acid in Seawater at Atmospheric Pressure. Limnol. Oceanogr. 18, 897-907.

Melo, L., Bott, T., 1997. Biofouling in water systems. Exp. Therm. Fluid. Sci. 14, 375-381.

Mitbavkar, S., Raghu, C., Pavan, D., 2012. Picophytoplankton community from tropical marine biofilms. J. Exp. Mar. Biol. Ecol. 426, 88-96.

Newbold, L.K., Oliver, A.E., Booth, T., Tiwari, B., DeSantis, T., Maguire, M., Andersen, G., van der Gast, C.J., Whiteley, A.S., 2012. The response of marine picoplankton to ocean acidification. Environ. Microbiol. 14, 2293-2307.

Oliver, A.E., Newbold, L.K., Whiteley, A.S., Gast, C.J., 2014. Marine bacterial communities are resistant to elevated carbon dioxide levels. Environ. Microbiol. Reports. 6, 574-582.

Parsons, T.R., Maita, Y., Lalli, C.M., 1984. Manual of chemical and biological methods for seawater analysis. Pergamon.

Patil, J.S., Anil, A.C., 2005. Quantification of diatoms in biofilms: standardisation of methods. Biofouling 21, 181-188.

Patil, J.S., Jagadeesan, V., 2011. Effect of chlorination on the development of marine biofilms dominated by diatoms. Biofouling 27, 241-254.

Qian, P.-Y., Lau, S., Dahms, H.-U., Dobretsov, S., Harder, T., 2007. Marine biofilms as mediators of colonization by marine macroorganisms: implications for antifouling and aquaculture. Mar. Biotechnol. 9, 399-410.

Riebesell, U., Fabry, V.J., Hansson, L., Gattuso, J.-P., 2010. Guide to best practices for ocean acidification research and data reporting. Publications Office of the European Union Luxembourg.

Rost, B., Zondervan, I., Wolf-Gladrow, D., 2008. Sensitivity of phytoplankton to future changes in ocean carbonate chemistry: current knowledge, contradictions and research directions. Mar. Ecol. Prog. Ser. 227, 227-237.

Russell, B.D., Connell, S.D., Findlay, H.S., Tait, K., Widdicombe, S., Mieszkowska, N., 2013. Ocean acidification and rising temperatures may increase biofilm primary productivity but decrease grazer consumption. Phil. Trans. R. Soc. B 368, 20120438.

Sabater, S., Admiraal, W., 2005. Periphyton as biological indicators in managed aquatic ecosystems. Periphyton. Ecology, exploitation and management. CABI Publishing, London, 159-178.

Siu, N., Apple, J.K., Moyer, C.L., 2014. The effects of ocean acidity and elevated temperature on bacterioplankton community structure and metabolism. Open J. Ecol. 4, 434-455.

Sommer, U., 1999. Periphyton architecture and susceptibility to grazing by periwinkles (*Littorina littorea*, Gastropoda). Int. Rev. Hydrobiol. 84, 197-204.

Stocker, T., Qin, D., Plattner, G., Tignor, M., Allen, S., Boschung, J., Nauels, A., Xia, Y., Bex, B., Midgley, B., 2013. IPCC, 2013: climate change 2013: the physical science basis. Contribution of working group I to the fifth assessment report of the intergovernmental panel on climate change.

Suzuki, Y., Takahashi, M., 1995. Growth responses of several diatom species isolated from various environments to temperature. J. Phycol. 31, 880-888.

Thomas, M.K., Kremer, C.T., Klausmeier, C.A., Litchman, E., 2012. A global pattern of thermal adaptation in marine phytoplankton. Science 338, 1085-1088.

Tolker-Nielsen, T., Molin, S., 2000. Spatial organization of microbial biofilm communities. Microb. Ecol. 40, 75-84.

Tomas, C.R., 1996. Identifying marine diatoms and dinoflagellates.

Torstensson, A., Chierici, M., Wulff, A., 2012. The influence of increased temperature and carbon dioxide levels on the benthic/sea ice diatom *Navicula directa*. Polar. Biol. 35, 205-214.

Torstensson, A., Hedblom, M., Andersson, J., Andersson, M., Wulff, A., 2013. Synergism between elevated pCO₂ and temperature on the Antarctic sea ice diatom *Nitzschia lecointei*. Biogeosciences 10, 6391-6401.

Toseland, A., Daines, S.J., Clark, J.R., Kirkham, A., Strauss, J., Uhlig, C., Lenton, T.M., Valentin, K., Pearson, G.A., Moulton, V., 2013. The impact of temperature on marine phytoplankton resource allocation and metabolism. Nat. Clim. Change 3, 979-984.

Trimborn, S., Wolf-Gladrow, D., Richter, K.-U., Rost, B., 2009. The effect of pCO₂ on carbon acquisition and intracellular assimilation in four marine diatoms. J. Exp. Mar. Biol. Ecol. 376, 26-36.

Verspagen, J.M., Van de Waal, D.B., Finke, J.F., Visser, P.M., Huisman, J., 2014. Contrasting effects of rising CO₂ on primary production and ecological stoichiometry at different nutrient levels. Ecol. Lett. 17, 951-960.

Worden, A.Z., Nolan, J.K., Palenik, B., 2004. Assessing the dynamics and ecology of marine picophytoplankton: the importance of the eukaryotic component. Limnol. Oceanogr. 49, 168-179.

Wu, Y., Gao, K., Riebesell, U., 2010. CO₂-induced seawater acidification affects physiological performance of the marine diatom *Phaeodactylum tricornutum*. Biogeosciences 7, 2915-2923.

Legends to figures

Fig. 1. Chlorophyll *a* concentration (μ g dm⁻²) in the biofilm grown under different pCO₂ (~400 μ atm and ~1500 μ atm), temperature (~30 °C and ~34 °C) and nutrient (unenriched and enriched) combinations. Values are mean ± SE (n=3). T = temperature, C = pCO₂, N = nutrients, + = elevated levels. TCN refers to control condition.

Fig. 2. Photosynthetic quantum yield (Fv/Fm) of the biofilm grown under different pCO₂ (~400 μ atm and ~1500 μ atm), temperature (~30 °C and ~34 °C) and nutrient (unenriched and enriched) combinations. Values are mean ± SE (n=3). T = temperature, C = pCO₂, N = nutrients, + = elevated levels. TCN refers to control condition.

Fig. 3. Abundance (cells dm⁻²) of (A) diatom, (B) picophytoperiphyton, and (C) bacteria in the biofilm grown under different pCO₂ (~400 µatm and ~1500 µatm), temperature (~30 °C and ~34 °C) and nutrient (unenriched and enriched) combinations. Values are mean \pm SE (n=3). T = temperature, C = pCO₂, N = nutrients, + = elevated levels. TCN refers to control condition.

Fig. 4. Relative abundance of different species of (A) diatom and (B) picophytoperiphyton in the biofilm grown under different pCO₂ (~400 µatm and ~1500 µatm), temperature (~30 °C and ~34 °C) and nutrient (unenriched and enriched) combinations. Values are mean \pm SE (n=3). T = temperature, C = pCO₂, N = nutrients, + = elevated levels. TCN refers to control condition.

Fig. 5. Carbon biomass (μ g dm⁻²) of (A) autotrophic periphytes, (B) bacteria, and (C) the ratio between bacterial carbon biomass and autotrophic carbon biomass (BB:AB ratio) in the biofilm grown under different pCO₂ (~400 µatm and ~1500 µatm), temperature (~30 °C and ~34 °C) and nutrient (unenriched and enriched) combinations. Values are mean ± SE (n=3). T = temperature, C = pCO₂, N = nutrients, + = elevated levels. TCN refers to control condition.

Fig. 6. Carbon biomass (μ g dm⁻²) and percentage carbon contribution to total autotrophic carbon biomass from (A) larger (> 2 μ m) phytoperiphytes and (B) picophytoperiphytes (< 2 μ m) in the biofilm grown under different pCO₂ (~ 400 μ atm and ~1500 μ atm), temperature (~30 °C and ~34 °C) and nutrient (unenriched and enriched) combinations. Values are mean ± SE (n=3). T = temperature, C = pCO₂, N = nutrients, + = elevated levels. TCN refers to control condition.







Fig. 3



Fig. 4



Fig. 5

